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SEP 18 2001

To: YMP\_SR@ymp.gov  
cc:

Subject: Report

**Part of Records Package / Supplement / Correction**

Jacob Paz wrote:

Attached is my final report comments on the suitability of YMP as a high nuclear repository. I will have a short ammdmet which will be sent in few days.



- The question to be asked.doc

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## 1. Introduction

The question should be asked should Yucca Mountain be approved as High Nuclear Repository? At the present time there are major uncertainties and insufficient scientific data, which has been ignored or has not been completely investigated by YMP. There is a definite need for additional investigation before YMP, could be approved as a high nuclear repository. My conclusions is based upon analysis and review of the Environmental Impeach Statement (EIS), Supplement to the EIS, the Science Engineering reports, the US Environmental Agency publications and the professional literatures. Here are my comments:

## 2. Yucca Mountain Project

Will Yucca Mountain Repository probably will become some at that time in the near future, a Resource Conservation and Recovery Act (RCRA) and finally a CERCLA site at that time the corrosion of canisters and engineering barriers containing heavy metals as well as possible escaping radioactivity needs to considered. While the YMP management has the duty to disclose and communicate this potential hazard to the public, the draft EIS and the supplement to the EIS had ignores it. The U.S. Environmental Protection Agency (EPA) has set very stringent requirements concerning the long-term risk from RCRA and CERCLA sites containing heavy metal carcinogens to ensure the safety of the public in perpetuity. These regulations stand in sharp contrast to the requirements contained inapplicable laws and regulations protect the public from the effects of a geologic repository containing nuclear waste for only 10,000 years. The long-term assessments of risks associated with heavy metals mixtures are lacking in the EIS and its supplements.

The land disposal restrictions and requirements set by the U.S. Environmental Protection Agency (40 CFR 268) for (RCRA) metal carcinogens currently specify stated that land disposal sites cannot be located in seismically active regions cited by Okrent and Xing, (1). They further noted that there is an inconsistency in current regulations and practices such as the approval of YMP as a high nuclear waste repository. Specifically, there is a very strong probability that YMP will not be in compliance with both RCRA act (40 CFR 268); and the Emergency Planning and Community Right to know acts (EPCRA) section 313(d)(2)(B) and (C). The questions to be e asked how YMP Finally, could YMP provide experimental data that indicated that metal of canisters and engineering barriers that they would not be corded in order to obtain an exemption from EPA under current regulations (2)? To end with, what is the EPA position and how they are going to enforced their own regulations and standards.

### 3. Sorption of Heavy Metals and Radionuclides

Upon reviewing the YMP EIS, the supplement to the EIS, and the YMP Science and Engineering Report, I have noted that the estimation of the releases rates of heavy metal and radionuclides from the proposed repository may be in error: It appears that YMP scientist had failed properly to investigate the effects of heavy metals mixtures such as; Ni, Co, Cr, and Mo; and radionuclide mixtures on Zeolite sorption rate, affinity, break-point, and the potential replacement and the release rates of heavy metal and radionuclides into the environment. There is a good probability of an increase over the estimates of heavy metals or radionuclides release into the biosphere: At the present time, the health risk pose to populations, as calculated by YMP remains uncertain and needs additional study. Further more, most of Zeolites sorption studies were carried in small laboratory batch and should be addressed by additional research using large columns, and field experiments. Major errors could result from scaling results up from lab experiments to field situations without adequate validation.

### 4. Chromium Oxidation

YMP has concluded that the canister corrosion by contact with rock, would promote the formation of Silicate can reduce  $\text{Cr}^{+6}$  to  $\text{Cr}^{+3}$  was reported by Eary and Rai (3); he further noted that the rate of reduction of  $\text{Cr}^{+6}$  is also influence by the organic matter and  $\text{HS}^-$ . Reduction of  $\text{Cr}^{+6}$  and by microorganisms under anaerobic conditions was been reported in the literature by Martin et al (4). Palmer and Puls (5); reported that the oxidation of  $\text{Cr}^{+3}$  to highly toxic  $\text{Cr}^{+6}$  is carried out by oxygen and Manganese dioxide ( $\text{MnO}_2$ ). Equation 1. illustrates the oxidation of  $\text{Cr}^{+3}$  to  $\text{Cr}^{+6}$  by  $\text{MnO}_2$ .



Additionally, chromium  $\text{Cr}^{+6}$  is a highly toxic and carcinogen species in oxidation state  $\text{Cr}^{+6}$  can be reduced by  $\text{Fe}^{+3}$  to less toxic species  $\text{Cr}^{+3}$  as shown

However, review of Appendix I. pp I-17 and TRW199ba, it appears that YMP-EIS did not investigate the oxidation of  $\text{Cr}^{+3}$  to  $\text{Cr}^{+6}$  by manganese oxidation in the unsaturated saturated zone and in appropriate aquifer should be further investigated. In spite of Zielinski (6) having reported the present of manganese oxide at various locations at NTS and YMP in large quantities. The EQ6 simulation computer model in the EIS Appendix I. did not clearly provide a clear data shows oxidation of  $\text{Cr}^{+3}$  to  $\text{Cr}^{+6}$  by  $\text{MnO}_2$ . The information reported remains uncertain needs further input and probably additional research. Therefore, the rate of chromium releases rate into the bios remains uncertain and needs further examination. There is a very good potential for elevated levels of  $\text{Cr}^{+6}$  in the unsaturated, saturated zone included the groundwater, and the potential increases of human health risk.

### 5. The Nevada Test Site Groundwater Contamination

The historic activities at NTS include atmospheric weapon testing, underground nuclear testing, safety testing of nuclear weapons, nuclear weapons development, and

the disposal of low levels of radioactive waste. From 1951 to 1992 more than 820 underground nuclear tests and 100 atmospheric tests were conducted at NTS. About 820 underground nuclear tests have had been conducted at the Nevada Test Site. Of these 259 tests are presumed to have an impact on groundwater. Of these 112 were detonated below the water table. Tritium is the radionuclide of major concern because of its transportation properties, Hechanova and Hodge (9). The reported a tritium inventory of 69.9 MCi in the Pahute Mesa region and 30.7 MCi in the other regions of NTS.

The total underground radiological contamination of all radionuclides is about 310 MCi essentially all from underground testing. However, the 112Mci underground radiological source term considered in the EIS as being available for potential migrations is just the total from all underground tests that were conducted beneath the water table or within 101 Meter of the top of the water table, and 90% of this is tritium DOE (10) and Croff (11). The toxic materials present after nuclear detonation occur in three locations: 1). Incorporated into the melted glass pools in the bottom of the cavity, 2). Deposited on the rubble and along fractured surfaces within and outside of the cavity, and 3). Finally, the gases that are escape into the atmosphere within a short time after detonation of a nuclear device. The distribution of radionuclides is complex, and their behavior or deposition is not well understood Smith (12).

There is considerable uncertainty concerning the actual quantity of radioactivity that can be mobilized by leaching of contaminated subsurface debris by groundwater. Smith et al., (13) have summarized the uncertainties associated with leaching for the NTS and concluded that the radionuclides most likely to become mobile and migrate via the groundwater regime are: (1) tritium; (2) a number of anions and neutral species such as Tc-99, Ru-106, Cl-36, and I-129, all assumed to migrate at the same rate as groundwater; and (3) cationic species, including Sr-90, Cs-137, Co-60, Zr-95, Pu-239, and others, that are believed to move more slowly than groundwater to varying degrees. It should be noted that Zr-95, and Ru-106, all have half-lives less than three years and are not likely to pose a groundwater hazard; the same is probably true for cobalt-60 with a half-life of 5.2 years. However, the quantitative estimates are highly uncertain to the point of being almost non-existent. There has been essentially no study of whether the substantial fraction of the radiological source term that was deposited above the water table is moving downward into the saturated zone Borg, et al.,(14); and Krsting et al., (15).

The situation related to retardation of radionuclides transportation and by sorption into or onto rocks is somewhat better known than for leaching, with several studies having been conducted. Tritium is appropriately assumed to move at the same rate as the groundwater. However, documentation for most other radionuclides indicates that retardation factors vary significantly with respect to water composition, experimental conditions, and rock type. The causes of the variations are speculative Smith, (13). In fact, Daniels, (16) assumed no sorption of any radionuclides because of the limited database. Insoluble or highly retarded radionuclides can be transported by forming or attaching to colloidal particles, which then move essentially at the same rate

as the groundwater in which they reside. Kersting, (17) concluded that a substantial fraction of radionuclides could be associated with colloids, but the effects on transportation are not known. Contaminant transport by non-radioactive organic chemicals or degradation products thereof has not been studied or taken into account.

Pahute Mesa, which is the location of most of the U.S. large nuclear explosions, contains approximately 70 percent of the tritium at the NTS. Modeling results also indicate that groundwater flow paths from Pahute Mesa are the shortest of all those at the NTS site and constitute the highest potential for contamination migration to off-site public receptors IT Corporation (18). From recent analysis of water from a well near the TYRO nuclear weapon test site on Pahute Mesa the experimental data show that Pu-239 seem to be immobilized in groundwater; however, tests of two wells near the TYBO underground nuclear test at Pahute Mesa, at the Nevada Test Site do not confirm this. Test results showed that presence of Pu-239 in association with colloids, found at significant levels in well number ER-20-5 #1 at a depth of 860 m. While, at well number ER-20-5 #3 30 m south of #1 only a very small amount of Pu-239 was detected,, Kersting et al.,(19). All of the Pu-239 detected was shown to be associated with colloidal particles.

The GeoTrans (20) carried out tests for tritium; the experimental data were far below 20,000 pCi/L., which is EPA's allowable tritium concentration in drinking water. The study reported by Daniels (15) predicted much higher values. The estimated range of peak tritium concentrations at the closest uncontrolled use area varies from  $5 \times 10^{-4}$  pCi/L (arriving 150 years after the beginning of migration) to 3,800 pCi/L (arriving in 25 to 94 years). The hypothetical maximally exposed individual at this location is estimated to have a lifetime probability of contracting a fatal cancer between  $8 \times 10^{-12}$  (about one in one trillion) and  $1 \times 10^{-5}$  (about one in 100,000), depending on which model is used. These estimates are self-characterized as being conservative. The results indicate that at the Area 20 (Pahute Mesa) boundary of the NTS and at Oasis Valley the lifetime committed effective dose for other radionuclides is about 10 percent of that from tritium. Important radionuclides other than tritium were Sr-90, I-129, Cs-137, Ra-226, Pu-239, and Am-241. The risks from toxic chemicals resulting from weapons tests have not been estimated.

Of the big major concern is that the facts that the Underground Test Area Program (UGTA) strategy does not utilize risk as a major factor in how and where the DOE applies its resources to protect human health from contaminated groundwater at the NTS. Since the DOE does not have enough data to define adequately the hydrologic source term, an acceptable risk assessment for the groundwater contamination cannot be properly developed at this point. The baseline risk assessment for the NTS groundwater contamination is described as incomplete since it only characterizes the radioactive isotope tritium. Both YMP and NTS risk assessment ignored the potential health effect of mixed irradiation and toxic chemicals required data are unavailable or uncertain and this matter must be further investigated.

The focus on tritium is logical because it enters the groundwater easily since it is an isotope of hydrogen, and it has the highest inventory of any radionuclides at the NTS. But other radionuclides may travel as conservatively as tritium, and not be retained in the aquifer materials contaminated by testing. Np-237, Tc-99 are thought to be isotopes that can simulate tritium-like migration. In fact, neptunium is the major long-term culprit predicted to carry contamination from Yucca Mountain to offsite, down gradient locations. Since the DOE does not know the concentration of all radionuclides in the groundwater from nuclear testing, it cannot conduct an acceptable risk assessment for UTGA problem. If the UGTA strategy were to incorporate risk as a driver in the quest to understand, locate, and protect human health from contaminated groundwater, then one must look to the northwestern section of the NTS called the Pahute Mesa area. Pahute Mesa is where the largest and deepest underground nuclear tests were conducted in the volcanic rock aquifers. Specifically, in the western Pahute Mesa area some shots were conducted so close to the NTS boundary that contamination may have been injected off the NTS and into U.S. Air Force lands. YMP management had failed to incorporate NTS risk assessment into YMP risk assessment; the projected cancer risk in the Draft EIS needs a major revision.

## **6. Risk Assessment Exposure of Complex Mixtures**

Both the Environmental Protection Agency (EPA) and the Nuclear Regulatory Commission (NRC) have proposed radiation standards for drinking water near the YMP. The EPA (21) had issued a radiation protection standard of 15-mRem effective dose per year for YMP. While the NRC proposed corresponding radiation standard of 25 mRem. In addition the EPA is set a drinking water standard of 4 mRem at the nearest accessible site to Yucca Mountain. The EPA applied a cancer risk factor ranging from  $10^{-6}$  to  $10^{-4}$ , to be consistent with the existing policy under the Comprehensive Environmental Response, Compensation and Liability Act; and more recently, the Food Quality Protection Act. The Food Quality Protection Act requires a cancer standard risk factor of no greater than  $10^{-4}$ .

The EPA acknowledged that most radioactive sites are also contaminated with non-radiological toxic chemicals, but they failed to take into account the potential synergistic or antagonistic interactions of toxic chemicals with radionuclides at low concentrations. Neither the Department of Energy (DOE) nor the NRC has regulations or policies to address the possible problem associated with chemical interactions with radionuclides. The NRC has proposed a protection standard of 25 mRem effective dose per year for YMP (22) based on an acceptable cancer risk of 1 in 1000.

Recently, there has been an increasing concern among regulatory agencies and the public over the exposure to and possible adverse effects from exposure to complex mixtures of environmental pollutants (toxic chemicals). The EPA in 1986 and in 1990 (23-24) recognized the importance of complex mixtures and issued guidelines for the risk assessment of complex mixtures. The National Research Council (NRC) in 1988 (25) addressed concerns regarding exposures to complex mixtures. The Presidential/Congressional Commission on Risk Assessment and Risk Management in

1977 (26) stated that it "considered the risk assessment of mixtures to be a matter of considerable concern and importance." Additionally, the National Council on Radiation Protection and Measurements (NCRP), in 1993 (25), specifically acknowledged that a gap exists between chemical and radiation risk estimate. In addition, the NCRP confirmed that further study is needed to address issues such as damage to the immune system, and possible combined effects of chemicals and irradiation causing either synergistic or antagonistic effects.

In addition, RCRA Section 3004(m), which is, requires EPA to "promulgate regulations specifying those levels or methods of treatment, if any, which substantially diminish the toxicity of the waste or substantially reduce the likelihood of migration of hazardous constituents from the waste". On January 14, 1986, EPA proposed an approach for developing treatment standards under 3004(m) using technology-based levels determined by the performance of Best Demonstrated Available Technologies (BDAT) in conjunction with risk-based standards (screening levels). After receiving extensive comment on the proposed rule, EPA chose to promulgate only the technology-based level or BDAT approach. The U.S. Court of Appeals for the D.C. Circuit upheld EPA's technology-based approach to LDR; Hazardous Waste Treatment Council vs. EPA, 886 F. 2d (D.C. Cir. 1989).

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## **7. Mixed Irradiation Risk Assessment Models**

Several models has been proposed for the simulated the action of mixed irradiation with two types of radiation have been proposed in the last two decades, but YMP management failed to include them in the EIS. Mixed irradiation is sometimes composed of more than two types of radiation, and for this type of mixed irradiation, no model has yet been proposed. It is of importance to assess the effect of mixed irradiation in terms of the environment, groundwater contamination, transportation accidents, space, and medicine. Theoretical models for mixed irradiation with two types of radiation have been presented by Zaider and Rossi (27); and by Scott (28) based on the Theory of Dual Radiation; Tobias et al., (29); Ager and Haynes (30); Lamb (31); Suzuki (32) also have analyzed the action of mixed irradiation using their own models. However, mixed irradiation is sometimes composed of more than two types of radiation.

Suzuki (33) has developed a model that can be applied to any type of mixed irradiation (i.e., any time-lag) with two types of radiation (i.e., the extended Zaider-Rossi model). He also stated "it is difficult to extend this to a model involving more than two types of radiation and to avoid this difficulty, limited the model to simultaneous irradiation.

In radiobiological studies, very low dose-rates are usually concomitant with very long irradiations and vice versa since the doses used are those that give rise to the changes to be determined (i.e., neither too high nor too low doses). There are no very long irradiations at high dose rates or very short irradiations at low dose-rates in studies. Therefore, the terms very low dose rate and very long irradiation have the same meaning. Though this model is limited to simultaneous irradiation, it would be useful for assessing the effects of such irradiation, because no model has been reported for mixed irradiation with multiple types of radiation and because mixed irradiation often occurs simultaneously in nature. The action of mixed irradiation must be further investigated, Suzuki (34).

## **8. RCRA AND LDR**

The 1984 Hazardous and Solid Waste Amendments (HSWA) to the Resource Conservation and Recovery Act (RCRA) restrict the land disposal of hazardous wastes, including mixed waste. This overview outlines the major aspects of the land disposal restrictions (LDR) as they apply to mixed wastes.

## **9. Variances from the Treatment Standards**

The EPA recognizes that some mixed waste might not be treatable by the method or to the level specified in such situations, EPA will allow petitions to be submitted requesting a variance from the treatment standard. If granted on a national basis, these variances result in the establishment of a new treatability group and new treatment standards for all wastes in the treat ability group. Variances may also be granted on a site-specific basis. Site-specific variances may be granted administratively (i.e., without notice-and comment rulemaking) and have no generic application to similar wastes generated at other sites. Variance petitions should be sent to the U.S. EPA

## **10. Treatment in Surface Impoundment Exemption**

Treatment of wastes that are normally prohibited from land disposal is allowed in a surface impoundment or a series of surface impoundments that meet the technological requirements of 40 CFR 268.4(a)(3). After treatment, if the residues do not meet the applicable treatment standard (or statutory prohibition level if the treatment standard has not been established), then the residues must be removed for subsequent management within a year of entry into the unit and may not be managed in another surface impoundment. Also, a certification that attests that the technical requirements are met and a modified waste analysis plan that incorporates 40 CFR 268.4 residual testing requirements must be sent to the Regional Administrator.

## **11. Dilution as Treatment**



Under the LDRs, dilution is prohibited as treatment for both listed and characteristic wastes (see 40 CFR 268.3). However, exceptions to the prohibition were made for:

1. Certain characteristic wastes generated and managed in waste treatment systems regulated by the Clean Water Act (See 40 CFR 268.3(b)). (Note that prohibited wastes treated by inappropriate methods are considered impermissibly diluted.)
2. Listed and characteristic wastes that are aggregated for legitimate treatment in centralized treatment systems. (Note that centralized treatment of incompatible waste streams is not considered legitimate treatment and is viewed as impermissible dilution.)
3. Characteristic wastes that are disposed into hazardous or non-hazardous Class I injection wells regulated under the Safe Drinking Water Act and do not exhibit any prohibited characteristic of hazardous waste at the point of injection.
4. Prohibited non-toxic ignitable, reactive and corrosive wastes that are treated by dilution to meet a treatment standard.

None of four sections is applied to YMP as a basis for issue a variance to YMP as Dilution as Treatment method for treating groundwater contaminated with radionuclides.

## 12. Conclusion

In conclusion, in spite of governmental, professional and quasi-governmental organization recommendations and publication, the YMP management, did not address adequately issues of complex mixtures in the EIS. This should have addressed complex mixtures including heavy metals found in C-22 canisters and engineering brayers (Ni, Cr, Co, and U); neutron posing elements (Gd); and radionuclides (Tc-99, I-129, Np-237, U234, Pu-239). The health cancer risk posed to population is unknown and should be investigated before YMP be approved as a high nuclear Repository. While the US Congress mandate that DOE should investigate the suitability of YMP as high Nuclear Repository, YMP took the following position it is regulatory agencies responsibilities to investigate the effects of complex mixtures and risk. However, they forgotten they are in charge of investigation the site and they wrote the risk assessment.

YMP risk assessment dose not fully and properly incorporate the UGTA into their risk assessment, which raises a very serious scientific question such as: what is the full impact of groundwater from NTS on YMP risk model for example the tritium plume. The only way to address these issues is by additional research. YMP and NTS managements must work together and develop appropriate risk assessment model based upon experimental data being investigated with PBPK model and complex mixtures testing; using the best available science as advocated by US President.

There is several uncertainties that must be further are investigated such as: 1). What is the impact of metal mixtures on radionuclides mixtures on sorption, affinity, competition, by Zeolite and the release rate into the environment by Zeolite; 2). The effect of manganese oxide on oxidation of  $\text{Cr}^{+3}$  to  $\text{Cr}^{+6}$ , and the levels of chromium discharged into the environment needed to be further studied; 3). Most of Zeolites sorption studies were carried in small laboratory batch and should be addressed by additional research using large columns, and field experiments. Major errors could result from scaling results up from lab experiments to field situations without adequate validation.

Both the YMP, EPA and the state of Nevada must solve the issues associated with YMP high-level nuclear site will become in the near future a RECR, later CERCLA site respectively, and finally a mixed waste site. YMP must comply with the letter of the law and all EPA regulations. Finally, what is EPA position on these matters and how they are going to enforce their own regulations and guidelines?

### 13. References

1. Okrent D., and Xing L., J. of Haz. Mat. 38:363-384, 1993
2. EPA -260-R-01-002 Reports on the Corrosion of certain Alloys, July 2001.
3. Yucca Mountain Project Environmental Impact Statement Draft Proposal Appendix I, pp. I. 16- 18, I. 18, August 1999.
4. Eary L.E., and Rai D., Kinetics of Chromium (III) Oxidation to Chromium (IV) by Reaction with Manganese Dioxide, Environ. Sci. and Tech. 21:1187-1193, 1987.
5. Martin C., Boone D.R., and Palmer C.D., Chromate Resistant Microbes from Contaminated Soil and their Potential Bioagumetation Reduction of Cr (VI). Proceeding of Eight National Outdoors Action Conference and Exposition, Minneapolis, MN May 23-25, 1994. National Ground Water Association: 191-204, 1994.
6. Palmer C.D., and Puls R.W., Natural Attenuation of hexavalent Chromium in Ground water and Soil, EPA/504/5-94/505, 1994.
7. Zielinski R.A., Evaluation of Ash-Flow as Hosts for radioactive Waste: Criteria Based on Selective Leaching for Manganese Oxide. USGS/Open-File Reports 83-480, 1983.
8. Carlos A.B., Bish D.L. and Chipera S.J., Manganese-Oxide Minerals in Fractures of the Center Flat Tuff in Drill Core USW G-4 Yucca Mountain, Nevada, Los Alamos, LA-11787-MS, UC—814, July, 1990.
9. Triay I.R., Meijer A., Conca J. L., Kung K.S., Rundberg R.S., Stietlemeire B. A., and Tail C. D., Summary and Synthesis Report on Radionuclides Retardation for the Yucca Mountain Site Characterization Project, LA-13262\_MS, March 1997.
10. Hechanova T., and Hodge V., Draft Sources of Term Screening of Underground Nuclear Detonations a Nevada Test Site Herry Reid Center for Environmental Studies Sept. 1998.
11. U.S. Department of Energy, United State Nuclear Tests, July 1945 to Septmeber 1996. Nevada operation Office of External Affair DOE/NV-209 (Review 14), 1996.

12. Croff A, C., Disposition of the Nevada Test Site, In: Long Term Institutional management of U.S. Department of Energy, pp141-148, Pub. National Academy of Science, 2000.
13. Smith D.K., A Review of Literature Pertaining to Leaching and Sorption of Radionuclides Associated with Nuclear Explosive Melt Glass. Lawrence Livermont Laboratory UCRL-ID-113370, Lawrence Calif. 1993.
14. Smith D. k., Kersting A.B., Rose J. M., Kennleally T.P., Hudson H.B., Eton C.F., and Davidson M., Hydrological Resources Management program and Underground Test Operation Unit FY 1997 Progress Report Lawrence Livermont Laboratory UCRL-ID-1130792, Lawrence Calif. 1998.
15. Borg L. Y., Rose H.E., Levy H.B., and Ramsptt L.D., Information Pertinent o Migration of Radionuclides in Groundwater at the Nevada Test Site-Part I. Review and Analysis of Existing Information Lawrence Livermont National Laboratory UCRL-ID-52078, Lawrence Calif. 1976.
16. Kersting A.B., Efurd D.W., Finnegan D.L., Rokop D.J., Smith D.K., and Thompson J.L., Nature397: 56-59, 1999.
17. Daniel J.I., Pilot Study Risk Assessment for Selected Problem at the Nevada Test Site. Livermont National Laboratory UCRL-ID-11389, Lawrence Calif. 1993.
18. Kersting A.B., The State of the hydrological Source Term Lawrence Livermont National Laboratory UCRL-ID-126557, Lawrence Calif. 1996.
19. IT Corporation. Corrective Action Unit Modeling Approach for tire Underground Test Area, Nevada Test Site, Nye County, Nevada Report DOE/NV/13052501, Las Vegas, N 1998.
20. GenTrans. Ins. A Fracture/Poxous Media Model of Tritium Transport in the Underground Weapons Testing Areas. Nevada Test Site. GooTrans, Ins., Boulder, Colo. 1995
21. Federal Register Part II. Environmental Protection Agency, 40CFR Part 197 Environmental Protection Radiation Standard for Yucca Mountain Project, Final Rule August 14, 2001.
22. Federal Registrar Part II. Nuclear Regulatory Agency, 10 CFR Part 19. Disposal of High -Level Radioactive Waste in a Proposed Geological Repository at Yucca Mountain, Nevada; Proposed rule, February 22, 1999.
23. EPA Guidelines for health Risk Assessment of Chemical Mixtures, Fed. Regist. 51, (185), 3414-24025, 1986.
24. EPA Risk Assessment Guideline for Superfund Vol. I. Part A. EPA/540/1-89/002
25. The National Research Council, Complex Mixtures, Methods for In Vivo Toxicity National Academy Press Washington. D.C.,1988.
26. The Presidential Congressional Commission on Risk Assessment and Risk Management 1997.
27. National Council on Radiation Protection Measurements (NCRP), Research needs for Radiation. NCRP ReportNo.117, NCRP Press, Bethesda, MD, 1993.
28. Zaider. M., and Rossi. H. H., Radiat. Res. 83: 732-739, 1980.
29. Scott. B. R., Bull. Math. Biol. 45: 323-3-16, 1983.

30. Tobias. C. A., Blakely. E. A., Ngo. F. Q. H., and Young, T. C. H., The repair-miss-repair model of cell survival. In: Radiation Biology in Cancer Research. Ed. Withers, H.R. M., pp. 195-230, Raven Press. New York, 1980.
31. Ager. D. D., and Haynes. R. H., Radiat. Res. 110: 129-1141, 1987.
32. Lam. G. K. Y., Radiat. Res. 110: 232 -243, 1987.
33. Suzuki. S., (1994). Radiat. Res.,138: 297-301, 1994
34. Suzuki. S., Environ. Health Perspec. Suppl. 105: I-155-1 158.
35. Suzuki S., Radiat. Res 39:215-221,1998.
36. Biokintetic Models for Radionuclides In; Cancer Risk Coefficients for Environmental Exposure to Radionuclides EPA 402-R-99-01, pp 145-156, 1999.